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Effect of influent flow rate distribution on the performance of step-feed biological nitrogen removal process

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Abstract

The step-feed anoxic/oxic biological nitrogen removal process has been proposed as an attractive alternative for the conventional biological nitrogen removal process for the purpose of enhanced nitrogen removal. For step-feed process, the biological nitrogen removal efficiency is a function of influent flow distribution. In this study, the effects of influent flow rate distribution on the performance of nitrogen removal process were investigated. The effects of influent flow rate distribution on COD, ammonia, total nitrogen removal efficiency, nitrification rate and sludge volume index value were also studied. According to the performance characteristic of the step-feed process, the concept of influent flow rate distribution ratio was firstly introduced. The maximum influent flow distribution ratios (λ_{max}) under the condition of different influent C/N ratios were determined, respectively, by trial and error method. The experimental results showed that high total nitrogen removal efficiency, higher than 95%, could be achieved under certain influent flow rate distribution ratio without internal nitrate cycle or addition of external carbon source. It was obvious that nitrification rate of each stage under different influent flow distributions decreased along with the decreasing of sludge loading and volume loading in each stage, and the degree of reduction rise gradually with the increasing of influent flow distribution ratio. © 2006 Elsevier B.V. All rights reserved.

Keywords: Activated sludge; Biological nitrogen removal; C/N ratio; Flow rate distribution; Removal efficiency; Step-feed process

1. Introduction

To protect lakes and other natural water from eutrophication, stringent nutrient level is set for the effluents from the wastewater treatment plants. Most of the plants recently employ biological process. Various biological nutrient removal processes such as pre-denitrification process (A/O), anaerobic/anoxic/aerobic process (A2/O), University of Cape Town (UCT), modified Bardenpho processes and Virginia Initiative Plant (VIP) were developed and widely applied [1–6]. These approaches will require additional energy for liquid circulation and addition of external carbon substrate for denitrification in anoxic zones. Further due to the growth of autotrophic nitrifying organisms in the aerobic tank, external addition of alkaline source is necessary to

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neutralize pH. As a result, the operational cost of these processes will increase significantly. To overcome these, the step-feed biological nitrogen removal (SFBNR) process has increasingly been proposed as an attractive alternative. Usually, this proposed process consist two or more stages of denitrification–nitrification reactors in series. For SFBNR process, the energy for internal recycle is not necessary. In addition the solids retention time (SRT) can increase because of suspended solids gradient along the reactors [7]. As a consequence of these features the process has been found to offer relevant advantages for both new and existing plants [8,9].

However, the optimum design and operation of the step-feed process is a difficult task because of step feeding of influent flow and complexity of reactor configuration. Volume ratios of anoxic and aerobic zone and wastewater fraction to be diverted from the inlet of the system are important parameters to be considered in the design of the step-feed process. Wastewater characteristics, especially influent C/N ratio, significantly affect the design and operation.

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During the last decade, many researchers have put much attention to this process and drawn many valuable conclusions. In the aspects of theoretical analysis [10], theoretically analyzed on nitrogen removal of the step-feed anoxic-oxic activated sludge process and its application for the optimal operation. Compared with the conventional denitrification-nitrification process, the step-feed process with four stages might reduce about 25% the total reactor volume [11,7]. Practically, when summarizing the conceptual approach and evaluating the operation of Riva plant (in Istanbul) [12], found that the number of stages influenced significantly the operation and nitrogen removal efficiency in step-feeding system. Compared with the conventional denitrification-nitrification process, the step-feed process with three stages might reduce about 20% the hydraulic retention time. The Newtown creek wastewater treatment plant, the largest plant in New York City, was reconstructed utilizing step-feed process for enhanced biological nitrogen removal in 1996. The operating results from January 1997 to June 1998 showed the BOD (biochemical oxygen demand), SS (suspended solids) and total nitrogen removal efficiency was 82-86.1%, 84.5-89.5% and 76-85%, respectively [13,9]. The Lethbridge wastewater treatment plant in Canada was retrofitted as five stages step-feed process for biological nutrient removal (BNR). The operating results showed the average mixed liquid suspended solids (MLSS) and treatment ability was higher than conventional BNR process. The average effluent ammonia and nitrate concentration was lower than 0.5 and 5 mg/L, respectively, in the whole year of 1999 [8].

The total nitrogen removal efficiency can be enhanced for conventional biological nitrogen removal process given better influent wastewater feeding mode. The previous studies results also showed that biological nitrogen removal efficiency is a function of influent flow distribution [14]. But there is no further reports about the effect of influent flow rate distribution on the performance of step-feed biological nitrogen removal process [15,11]. In this study, this effect will be evaluated. Moreover, than this, the effects of influent flow rate distribution on ammonia and total nitrogen removal efficiency, nitrification rate and sludge volume index value will also be examined.

2. Material and methods

2.1. Reactor system

A four stages step-feed biological nitrogen removal process made of plexiglass with a working volume of 80 L was used in this study (Fig. 1). The two channels cuboid reactor has a dimension of $80 \text{ cm} \times 22 \text{ cm} \times 50 \text{ cm}$. Each stage consisted of an anoxic and an aerobic zone, and the ratio of anoxic reactor volume and aerobic was maintained at 1:3.5 for four same stages in this study. For maintaining plug-flow purpose and concentration gradient, the aerobic zones in each stage are separated by clapboards as three joint compartments. A mechanical mixer was used in anoxic zone to provide liquid mix well. A number of outlets for samples were placed at a distance of 20 cm from reactor bottom in each anoxic and aerobic compartment. An air compressor was used for aeration. An air flow meter was used for controlling the airflow rate in reactor. The type of final clarifiers is a vertical clarifier with working volume of 30 L.

2.2. Wastewater composition

The reactor feed consisted of synthetic wastewater with characteristics similar to those of domestic wastewater [16]. It was prepared by using tap water, dechlorinated by the use of sodium thiosulfate, and the addition of chemicals as indicated in Table 1. A few of glucose and amylum that are not very easily biodegradable organic materials were also added to supplement COD. In addition some other organic materials such as glucose and maltose also exist in the brewery wastewater. Nitrogen and phosphorus were adjusted by adding NH₄Cl and KH₂PO₄ to the feed water. Sodium bicarbonate was also added to adjust alkalinity. The wastewater was continuously fed to the reactor and the flow rates were controlled by four peristaltic pumps (Model ESBN4, Iwaka Cop. Japan), respectively.

2.3. Experimental operating procedure

The start-up of the SFBNR process was initiated by seeding the synthetic wastewater with the sludge came from the secondary clarifier of Harbin Wenchang wastewater treatment plant (A/O process, 100,000 m³/day), while the reactor was operated for 12 days in a batch mode to provide the initial colonization and accumulation of microorganisms. The reactor was then operated in a continuous flow mode by gradually increasing flow to promote bacterial growth. Steady state was reached after 28 days of operation. Once the steady-state was realized, various experiments were conducted under a total of eight different C/N ratios in which total nitrogen concentrations were maintained constant at 42 mg/L (40 mg/L of kjeldahl nitrogen concentrations) so as to compare the effluent total nitrogen concentration and total nitrogen removal efficiency.

During the experimental period, the SRT was controlled at 18 days using hydraulic control approach. The sludge returns ratio was set at 50 percent of influent flow rate controlled by a peristaltic pump (Model Z1515-18, Lange Bump Cop, China). The temperature of the reactor was kept at 22 ± 1 °C by temperature controller.

Table 1			
Compositi	on of syntl	hetic was	tewate

Compound	Concentration (mg/L)		
Brewery wastewater	9500-10500		
Glucose	500-650		
Starch	200-250		
NH ₄ Cl	114.6		
NaHCO ₃	900.0		
MgSO ₄ ·7H ₂ O	150.0		
NaCl	110.0		
CaCl ₂	85.5		
ZnSO ₄	90.0		



 influent tank 2. check valve 3. feed pump 4. mechanical mixer 5. diffuser 6. air flow meter 7. air compressor 8. secondary clarifier 9. effluent 10. return sludge pump 11. waste sludge

Fig. 1. Schematic diagram of step-feed biological nitrogen removal process.

Table 2

2.4. Samples and analytical procedures

The parameters measured including pH, temperature, DO, COD, NH₄⁺–N, TKN, NO₂⁻–N, NO₃⁻–N, MLSS, SVI, TN and alkalinity. Samples were prepared by filtering with 0.45 um Whatman filter papers. The measurements of DO, pH and temperature of wastewater were conducted daily using WTW pH/Oxi 340i (made in Germany). The type of probe of DO and pH was WTW CellOx 325 and pH-Electrode SenTix 4, respectively. All analyses were performed according to the Standard Methods [17].

3. Results and discussions

3.1. Introduction of influent flow distribution ratio (λ)

For this proposed process, theoretically, the biological nitrogen removal efficiency (η) for SFBNR process can be calculated by the following equation [7]:

$$\eta = \left(1 - \frac{\alpha}{1+R}\right) \times 100\% \tag{1}$$

where α is the ratio of flow rate distribution into the last stage to total influent flow rate and *R* is the sludge return ratio.

The concentration of total nitrogen in the effluent is determined by the kjeldahl nitrogen in the aerobic zone of the last stage, only if in each stage, complete nitrification and denitrification in aerobic and anoxic zone are realized, respectively. In the last stage, where there is no nitrate accumulation and the inflow is at the minimum ratio, total nitrogen concentration in the effluent can maintain at the lowest level.

To promise the nitrate formed in each stage to be denitrified completely, the carbon source is provided by the influent of the next stage. Under certain influent C/N ratio, there must be a maximum influent flow distribution ratio (λ_{max}) between the last two stages, which is critical for lowest effluent total nitrogen concentration. Universally, this ratio value ($\lambda_{1,2,...}$) exists between the first and second stage, the second and third stage, and so on. In theory, this ratio value (λ_i , i = 1, 2, ..., max) would be equal if the reaction is completely for every stage. However, for step-feed process, different sludge concentrations occur because

The maximum influent flow rate distribution ratio under different influent C/N ratios

Influent C/N ratio	Maximum influent flow distribution ratio (λ_{max})
6.75	1.75
8	2
9.25	2.25
10.5	2.5
11.75	2.75
13	3
15	3.5
17	4

the returned sludge will be unevenly distributed by the step-feed wastewater in each stage. So, assimilation varies even between two contiguous stages. In practice, the λ_i between any two stages will be larger than that between backward contiguous two stages. For example, $\lambda_1 > \lambda_2 > \cdots > \lambda_{max}$. For a process with four or five stages, there would be three or four λ_i . It would be tedious to get the each λ_i through checking all the values. But this ratio should be around the value of λ_{max} , especially when the λ_{max} is large. From the standpoint of process control, optimal operation and management, a uniform distribution ratio (λ_{max}) can be introduced in the process of influent flow rate splitting.

The influent flow distribution ratio (λ_{max}) is determined by the wastewater influent C/N ratio, and the value is unique under certain influent C/N ratio. The value of λ_{max} should be ascertained through the experimental methodology by trial and error

Table 3	
Percent of NH4 ⁺ -N removal	

λ	Percent removal (%)						
	Stage 1	Stage 2	Stage 3	Stage 4	Overall		
1.75	72.35	73.60	92.75	100.0	100.0		
2	69.37	71.37	92.25	100.0	100.0		
2.25	66.67	69.23	91.67	100.0	100.0		
2.5	65.79	69.55	86.96	100.0	100.0		
2.75	65.79	69.55	86.96	100.0	100.0		
3	62.39	65.57	76.92	100.0	100.0		
3.5	60.14	73.02	82.61	100.0	100.0		
4	59.80	78.13	95.41	100.0	100.0		



Q (flow rate) in (Lh⁻¹), HRT (hydraulic retention time) in (hrs), MLSS (mixed liquid suspended solids) in (mgL⁻¹), VLR (volumetric loading rate) in (kgCOD (m³.d)⁻¹), SLR (sludge rate loading) in (kgCOD(kgMLSS.d)⁻¹), and HLR (hydraulic rate loading) in (m³ (m².d)⁻¹)

Fig. 2. Summary of steady-state operating conditions for various influent flow rate distribution ratios.



Fig. 3. Stage effluent parametric concentration for various influent flow distribution ratios. In the *X*-axis alphabetic number A means anoxic zone, O means oxic zone, and the Arabic numerals mean the stage and the compartment, respectively: (\blacksquare) NH₄⁺–N; (\blacktriangle) NO_X–N.

method, which is the tedious and time-wasting working. Once the value of λ_{max} is determined, the influent flow rate in each stage can be calculated by the following equation:

$$Q_{\text{Total}} = \lambda_{\max}^{n-1} X + \lambda_{\max}^{n-2} X + \dots + \lambda_{\max} X + X$$
(2)

In which *n* is the number of stage and *X* is the influent flow rate distributing in the last stage.

Table	4
Stage	effluent COD concentrations

3.2. The effect of influent flow distribution on the performance of step-feed biological nitrogen removal process

The maximum influent flow rate distribution ratios (λ_{max}) were drawn under the different influent C/N ratios according to the principle mentioned above by the trial and error method. The experimental results for maximum influent flow rate distribution ratios were shown in Table 2. The steady-state operating con-

λ	Stage 1 (mg/L)		Stage 2 (mg/L)		Stage 3 (mg/L)		Stage 4 (mg/L)	
	Anoxic effluent	Aerobic effluent						
1.75	100	36	57	26	55	24	38	16
2	123	36	58	28	56	26	36	14
2.25	144	48	62	36	54	30	40	16
2.5	168	56	62	36	54	24	38	16
2.75	204	72	84	38	56	26	36	16
3	252	80	104	42	56	28	34	14
3.5	288	96	112	48	46	24	32	12
4	324	106	124	56	46	24	30	11



Fig. 4. Variations of DO concentration in each reactor under for various influent flow distribution ratios.

ditions for various maximum influents flow distribution ratios were shown in Fig. 2. With certain influent C/N ratio the maximum influent flow distribution ratio in the process of influent flow rate splitting was definite. Similarly, the resulting effluent NH_4^+ –N and NO_X –N concentrations were shown in Fig. 3.

In this study, numbered cases to adjust the ratio of anoxic and aerobic reactor volumes of every stage were also made. As prementioned in step-feed process the concentration of total nitrogen in the effluent can be determined by the kjeldahl nitrogen in the aerobic zone of the last stage if complete nitrification and denitrification in aerobic and anoxic zone are accomplished in each stage. To achieve complete nitrification in each stage, volume adjusting will bring some effects, but it will not improve significantly tot-N removal efficiency. Under certain influent flow distribution ratio, changing the volume of anoxic zone will not enhance the extent of denitrification. For example, when influent C/N ratio is 13, the maximum influent flow distribution ratio can only reach 3. So whether the denitrification performs completely or not is not only decided by the volume of the anoxic zone, but also mainly by the influent COD rate. At this C/N ratio, the influent flow distribution ratio will not be changed through enlarging the volume of anoxic zone. For real domestic or municipal wastewater, increasing HRT of each anoxic phase by adjusting volume ratio will give chance of slowly biodegradable COD converted into readily biodegradable COD [18,19]. But in this paper this effect is



Fig. 5. Effluent total nitrogen concentration and total nitrogen removal efficiency under certain influent flow distribution ratios. In figure, the abscissa denotes the maximum influent flow distribution ratios (λ_{max}) under the condition of different influent C/N ratios. The main ordinate denotes the total nitrogen concentration when operating under the condition of the maximum influent flow distribution ratios. The sub-ordinate axis denotes the total nitrogen removal efficiency accordingly.

insignificant because the synthetic wastewater is prepared with brewery wastewater and a few of glucose and amylum that are not very easily biodegradable organic materials were also added.

In addition, if influent carbon is enough and complete nitrification in each stage, the total nitrogen removal efficiency under the influent flow distribution ratio of 1:3:3:3 is lower than the value obtained from 1:5:3:1, no matte how to adjust the ratio of anoxic and aerobic zone volumes.

3.3. The effect of influent flow distribution on the biological nitrogen removal

The percent of NH_4^+ –N removal and effluent COD concentrations in each stage were shown in Tables 3 and 4, respectively. One hundred percent of NH_4^+ –N removal efficiency in the fourth stage under various influent flow distribution ratios was achieved although the hydraulic loading rate was high. From Tables 3 and 4 and Fig. 2 it was also noted that the NH_4^+ –N

λ=1.75

0.10

0.09

0.08

0.07

0.06

0.05

0.04

0.03

0.02

0.01

0.00

0.10

Nitrification rate (kgNH₄/kgMLSS.d)

0.10

0.09

0.08

0.07

0.06

0.05

0.04

0.03

0.02

0.01

0.00

0.10

removal efficiency decreased along with the increasing of volumetric loading rate and sludge loading rate, accordingly to the results of [20–22]. However, a marked difference was noted that from 58.66% (C/N ratio = 17) to 65.79% (C/N ratio = 9.25) of NH₄⁺–N removal efficiency could be achieved in the first stage despite the volumetric rate loading and sludge rate loading was high, especially under high influent flow distribution ratios condition, which were opposite to the results of [23–25]. However, these were attributed to the presence of simultaneous nitrification and denitrification.

In step-feed biological nitrogen process, the sludge returns to the first stage of the reactor. Fig. 2 illustrates the step distribution of MLSS concentration. The MLSS concentration will reach 3900 mg/L in the first stage. Higher MLSS concentration is benefit to simultaneous nitrification and denitrification [26,27]. Moreover, DO concentration is a very critical factor to affect SND. In this experiment the DO concentrations under different influent flow distribution ratios have also been studied, and shown in Fig. 4. When the process is operated with a

λ=2.25

0.10

0.09

0.08

0.07

0.06

0.05

0.04

0.03

0.02

0.01

0.00

0.10

λ=2





Fig. 7. Variation of sludge volume index under different influent flow rate distribution ratios.

certain influent flow rate ratio, DO is severely deficient in some stage. Those factors attribute to the occurrence of SND [28–30]. During recent years microbiologists have shown that nitrifying as well as denitrifying are of greater physiological variety than expected. Certain species of bacteria such as *Nitrosomonas*

europea and *Nitrosomonas eutropha* are able to denitrify aerobically [31–33]. However, many heterotrophic organisms have been found to be able to nitrify organic and inorganic nitrogen compounds [34,35]. Details about simultaneous nitrification and denitrification would be discussed in another article [36].

Effluent total nitrogen concentration and removal efficiency under certain influent flow distribution ratios were shown in Fig. 5. From Fig. 5 it should be emphasized that high total nitrogen removal efficiency, higher than 95%, could be obtained in step-feed biological nitrogen removal process under certain influent flow distribution ratio. When influent flow rate distribution ratio increased from 1.75 to 4, the total nitrogen removal efficiency could increase from 90.2% to 96.5%. But for conventional pre-denitrification process, 1850% internal recycle ratio is needed as well as 50% sludge recycle ratio. In theory, the tot-N removal efficiency formula of pre-denitrification process is described as Eq. (3) [4]:

$$\eta = \frac{R+r}{1+R+r} \tag{3}$$

in which *R* is the internal recycle ratio, *r* the sludge return ratio and η is the removal efficiency

Assuming *r* equals to 50% and η equals to 95%, we can get the value of *R* is 1850%. Because of microorganism assimilation and SND, it would be below this value in practical. But it is



Fig. 8. Typical microbial photograph under the conditions of different flow rate distribution ratios. Left: $\lambda = 1.75$; right: $\lambda = 2.75$; bottom: $\lambda = 4$. When $\lambda = 1.75$ or 2.75, the sludge was sampled to take picture after dyeing with Loeffler. When $\lambda = 4$, the sludge was taken to take picture without dyeing. The magnification is $40 \times 10 = 400$ times.

obvious that step-feed process has distinct advantage over the pre-denitrification process in saving operation costs.

The nitrification rates of each stage under different influent flow distribution ratios were calculated and shown in Fig. 6. From the figure the nitrification rates were all higher than the reported value of conventional biological nitrification rate [37], indicating the high efficiency of the step-feed process. Combined Figs. 2 and 6, it was obvious that nitrification rate decreased along with the decreasing of sludge loading and volume loading rate in each stage, especially when the process was operated at higher influent flow distribution ratio. The degree of reduction rose gradually with the increasing of influent flow distribution ratio. This would be significant for the process operation practically. At the cases of higher influent ammonia concentration, relative large part of influent flow should be attributed to the former stage so as to maintain the final effluent ammonia concentration at a low level.

3.4. Effect of influent flow distribution on sludge volume index

The sludge volumetric index (SVI) was also examined during the experimental period. After the process operating under a maximum influent flow distribution ratio (λ_{max}) and reaching in a steady condition, the sludge was sampled to measure the SVI value and make images of microorganism. The SVI value increased along with the increasing of influent flow rate distribution ratio, shown in Fig. 7. There were lots of filamentous bacterial in activated sludge under microscope (Fig. 8). The main reason for filamentous bacteria sludge bulking could be attributed to the high organic loading rate in the first stage and inefficient DO concentration. When influent flow rate distribution ratio was 3.5, for example, the concentration of DO in the first compartment of aerobic zone of the first stage was only 0.07 mg/L. But the SVI value could be returned to normal value after adjusting the influent flow distribution and feeding averagely. From the standpoint of loading equilibration and preventing filamentous bacteria sludge bulking, the relative average influent flow distribution should be maintained for certain TN removal efficiency.

4. Conclusions

The laboratory pilot scale studies were conducted to evaluate the effect of influent flow rate distribution on the performance of step-feed biological nitrogen removal process. The results of the study demonstrated that improvements in the total nitrogen removal efficiency of the step-feed biological nitrogen removal process could be obtained by adjusting influent flow rate distribution under certain C/N ratio. In particular the following conclusions are arrived at:

(1) The influent flow rate distribution ratio was firstly introduced in the step-feed process based on the performance characteristic. Under the condition of different influent C/N ratios of 9.25, 10.5, 11.75, 13, 15 and 17, the maximum influent flow distribution ratio (λ_{max}) was 1.75, 2, 2.25, 2.5, 2.75, 3, 3.5 and 4, respectively, according to the principle by the trial and error methods, and the total nitrogen removal efficiency was 90%, 92.075%, 93.625%, 94.6%, 95.4%, 95.95%, 96.3% and 96.5%, respectively, when there was no accumulation of nitrate in anoxic zone of the last stage and the minimum influent flow rate in the last stage.

- (2) The total nitrogen removal efficiency of the step-feed biological nitrogen removal process was a function of influent flow rate distribution ratio, whereas the maximum influent flow distribution ratio (λ_{max}) was determined by the influent C/N ratio.
- (3) Relative high nitrification rate, higher than 0.085 kg NH₄/(kg MLSS day) was achieved in the step-feed process. It was obvious that the nitrification rate of each stage decreased along with the decreasing of sludge loading and volume loading in each stage, especially when the process was operated at higher influent flow distribution ratio. The degree of reduction rose gradually with the increasing of influent flow distribution ratio. At the cases of higher influent flow should be attributed to the former stage so as to maintain the final effluent ammonia concentration at low level.
- (4) The sludge volume index value would also increase along with the increasing of influent flow distribution ratio. From the standpoint of loading equilibration and preventing filamentous bacteria sludge bulking, the relative lower influent flow distribution ratio should be adopted for certain TN removal efficiency.

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